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Abstract: Human biomonitoring provides an integrated measure for individual exposure to environmental pollutants. Better insight in inter-individual variability of biomarkers of exposure may help in the interpretation of biomonitoring studies.

The aim was to study the impact of outliers, determine the optimal unit for fat-soluble biomarkers in serum and quantify the major determinants for biomarkers of exposure to polychlorinated aromatic hydrocarbons (PCAHs) in three age groups.

Data were obtained from the Flemish Environment and Health Study (2002-06). Marker PCBs (sum of 138, 153, 180), hexachlorobenzene (HCB) and p,p'-DDE were measured in cord blood samples of 1196 newborns, in serum samples of 1679 adolescents (14-15 years) and 1583 adults (50-65 years).

Exclusion of influential outliers in multiple linear regression models lead to models that are better applicable to the general population. In terms of adjusted R-square, the regression model with the pollutant expressed in volume-based units and blood fat as a separate independent variable was superior compared to models with other units. We found highly consistent relationships between the serum concentration of PCAHs and blood fat, age, changes in body weight, animal fat in the diet, local vegetable consumption (HCB and p,p'-DDE only) and being breastfed as a baby (in adolescents only). The impact of sex and BMI differed by age. For biomarkers of persistent pollutants that reflect long-term exposure, the relation between the covariates and the biomarkers can be well quantified.

Mol, August 19, 2008

Environmental Research  
To the Editor

Dear Editor,

Please find enclosed a manuscript entitled 'Determinants of polychlorinated aromatic hydrocarbons in serum in three age classes – methodological implications for biomonitoring' by E. Den Hond, E. Govarts, L. Bruckers and G. Schoeters which we would like to submit for publication in Environmental Research.

The manuscript presents results of the Flemish Environment and Health Study, a large biomonitoring trial (over 4500 participants) that was conducted in three age classes. The analysis of the data has allowed us to deduct general guidelines for biomonitoring and advices for the design and analysis of future biomonitoring trials.

We hope that the manuscript can be considered for publication in Environmental Research.

Yours sincerely,

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DETERMINANTS OF POLYCHLORINATED AROMATIC  
HYDROCARBONS IN SERUM IN THREE AGE CLASSES -  
METHODOLOGICAL IMPLICATIONS FOR BIOMONITORING.

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1 **ABSTRACT**

2 Human biomonitoring provides an integrated measure for individual exposure to  
3 environmental pollutants. Better insight in inter-individual variability of biomarkers of  
4 exposure may help in the interpretation of biomonitoring studies.

5 The aim was to study the impact of outliers, determine the optimal unit for fat-soluble  
6 biomarkers in serum and quantify the major determinants for biomarkers of exposure  
7 to polychlorinated aromatic hydrocarbons (PCAHS) in three age groups.

8 Data were obtained from the Flemish Environment and Health Study (2002-06).

9 Marker PCBs (sum of 138, 153, 180), hexachlorobenzene (HCB) and p,p'-DDE were  
10 measured in cord blood samples of 1196 newborns, in serum samples of 1679  
11 adolescents (14-15 years) and 1583 adults (50-65 years).

12 Exclusion of influential outliers in multiple linear regression models lead to models  
13 that are better applicable to the general population. In terms of adjusted R-square,  
14 the regression model with the pollutant expressed in volume-based units and blood  
15 fat as a separate independent variable was superior compared to models with other  
16 units. We found highly consistent relationships between the serum concentration of  
17 PCAHS and blood fat, age, changes in body weight, animal fat in the diet, local  
18 vegetable consumption (HCB and p,p'-DDE only) and being breastfed as a baby (in  
19 adolescents only). The impact of sex and BMI differed by age.

20 For biomarkers of persistent pollutants that reflect long-term exposure, the relation  
21 between the covariates and the biomarkers can be well quantified.

22

23 **KEY WORDS**

24 biomonitoring – biomarkers of exposure – polychlorinated aromatic hydrocarbons  
25 (PCAHS) – pesticides – polychlorinated biphenyls (PCBs)

1 **FUNDING SOURCES**

2 This study was supported by a grant of the Long Range Research Initiative (LRI) of  
3 the European Chemical Industry Council Cefic (HETRA D2.2).

4

5 The database was obtained with permission from the Flemish Environment and  
6 Health Study Group. The Flemish Environment and Health Study was commissioned,  
7 financed and steered by the Flemish Community (department of Science, department  
8 of Public Health and department of Environment), without any responsibility for the  
9 scientific content.

10

11

12 **APPROVAL**

13 The study was approved by the medical ethical committee of the University of  
14 Antwerp (July 4, 2002).

15

# 1 INTRODUCTION

2  
3 Biologic monitoring, i.e. biomonitoring, is used to assess human exposures to  
4 environmental and workplace chemicals. Biomarkers of exposure take into account  
5 inter-individual differences in absorption, distribution, biotransformation, and  
6 excretion of a substance which may be associated with differences in age, sex,  
7 height, weight, physiologic and nutritional status, duration of exposure, etc. Hence,  
8 biomarkers of exposure assess the internal dose of a xenobiotic compound and are  
9 therefore likely to be directly associated with possible adverse health effects  
10 (Lauwerys and Hoet 2001).

11 Between 2002 and 2006, a large biomonitoring campaign was conducted in  
12 Flanders, the Dutch-speaking part of Belgium. As biomarkers of exposure to  
13 polychlorinated aromatic hydrocarbons (PCHAhs), we studied the serum concentration  
14 of polychlorinated biphenyls (sum of marker PCB 138, 153 and 180) and serum  
15 concentrations of the organochlorine pesticides hexachlorobenzene (HCB) and p,p'-  
16 dichlorodiphenyldichloroethylene (p,p'-DDE) in more than 4400 participants of the  
17 general population including newborns (n=1196), 14- to 15-year old teenagers  
18 (n=1679) and adults between 50 and 65 years (n=1583).

19 The aims of this study were to use the data of this large cohort to investigate in detail  
20 the impact of outliers on summary outcomes and regression models, to compare the  
21 use of fat-adjusted *versus* volume-based units, and to determine the major factors of  
22 inter-individual variability of serum PCBs, HCB en p,p'-DDE in the three age groups.

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24  
25

# 1 MATERIALS AND METHODS

## 2 *Environment and Health Study*

3 Between 2002 and 2006 we have introduced in Flanders a biomonitoring network  
4 which includes measurement of biomarkers of exposure in combination with  
5 biomarkers of effect and individual health data. Flanders is an industrialized region in  
6 the North of Belgium with 6 million inhabitants, it is densely populated and has a  
7 dense traffic network. Participants were systematically recruited in eight geographical  
8 areas covering 22% of the surface area of Flanders and 20% of the population. Two  
9 areas were urbanized (Antwerp city and Ghent city), four areas were characterized  
10 by industrial settings (Ghent and Antwerp harbor, non-ferro industry, chemical  
11 industry and areas around waste incinerators), one area had intensive fruit cultivation  
12 (fruit growing area) and one area was less densely populated and had no registered  
13 emissions (rural area). Three age groups were involved in the human biomonitoring  
14 study: newborns and their mothers, 14-15 year old adolescents and adults between  
15 50 and 65. In total, about 4500 participants were systematically recruited. The  
16 recruitment campaign was scheduled over three years (2002-2004). Inclusion criteria  
17 for participation in the campaign were residing at least five years in the area, giving  
18 written informed consent and being able to fill in an extensive Dutch questionnaire.  
19 Participants were enrolled at random within primary sampling units. By stratified  
20 sampling by study area, we selected 26 maternities, 42 secondary schools and 46  
21 communities as primary sampling units for the newborn, adolescent and adult study,  
22 respectively. In the newborn study, 97% of the eligible mothers agreed to deliver cord  
23 blood and answer the questionnaire. In the adolescent and adult study, invitation  
24 letters were sent via the schools and by regular post. 71.6% of the adolescents and  
25 47.5% of the adults replied to the invitation and respectively 85.7% and 75.3% of



1 those who answered, agreed to participate. The study was approved by the medical  
2 ethical committee of the University of Antwerp.

3

#### 4 *Blood collection*

5 In newborns, minimal 30 mL cord blood was collected by leaving the umbilical cord  
6 blood running off in 50 mL polypropylene tubes filled with 0.5 mL Na<sub>2</sub>EDTA as  
7 anticoagulant. Cord blood was aliquoted and plasma was separated by centrifugation  
8 within 24h in the maternity laboratories. The aliquoted samples were kept in the  
9 refrigerator and transported to the analytical laboratory within one week. After  
10 transport in cool boxes, the samples were stored at -20°C until analysis.

11 Both in the adolescents and the adults, 40 mL of non-fasting peripheral blood was  
12 sampled. Na<sub>2</sub>EDTA (10% v/v) was added to whole blood while serum was prepared  
13 by immediate centrifugation of the coagulated blood. Samples were kept in cool  
14 boxes and transported to the central laboratory within 24h where they were stored at  
15 -20°C until analysis.

16

#### 17 *Measurements in cord blood*

18 Marker PCB 138, 153, 180 and the chlorinated pesticides HCB (hexachlorobenzene)  
19 and p,p'-DDE (p,p'-dichlorodiphenyldichloroethylene) were analyzed by gas  
20 chromatography equipped with an electron capture detector using the method of  
21 Gomara et al. (Gomara et al. 2002). In all three biomonitoring campaigns, chemical  
22 analyses were performed by two labs. Both laboratories participated in the AMAP  
23 proficiency testing scheme (Institut National de Santé Publique, Quebec, Canada).  
24 The measurement of uncertainty was estimated from the results of AMAP samples,  
25 and ranged between 21% and 34% for all the compounds except for HCB (64%). The

1 limit of detection for all chlorinated compounds in plasma was 0.02 µg/L. The plasma  
2 total lipid concentration was determined gravimetrically. In case no value could be  
3 obtained gravimetrically, total lipid concentration was calculated on the basis of  
4 routinely measured triglycerides and total cholesterol by the following formula: total  
5 lipids = 1.33 x (triglycerides + cholesterol) + 50.5 mg/dL (Covaci et al. 2006).

6

### 7 *Questionnaire data*

8 All participants completed an extensive questionnaire (self-administered), assessing  
9 information on lifestyle, dietary intake, use of tobacco and alcohol, residence history,  
10 health, education, hobbies and occupation (if applicable). Body mass index (BMI) of  
11 the mothers was calculated, based on self-reported height and weight before  
12 pregnancy. For adolescents and adults, height and body weight were measured  
13 according to a standardized protocol (WHO 1995). Dietary intake was assessed via a  
14 semi-quantitative food frequency questionnaire (FFQ) as described in detail by Bilau  
15 et al. (Bilau et al. 2008). For each individual, dietary intake of fat from different  
16 sources was estimated in grams per day, for different sources of fat i.e. beef, pork,  
17 sheep, horse, chicken, turkey, cereals, yoghurt, milk, eggs, cheese, cooking and  
18 frying fats, seafood (shrimps and mussels) and fish (lean, fatty, smoked and canned).  
19 Participants were asked whether they regularly use local food products (meat or  
20 vegetables); these answers were used as binary variables in the analysis.

21

### 22 *Statistical Analysis*

23 Database management and statistical analyses were done with SAS software  
24 version 9.1 (SAS Institute Inc.).

1 Biomarkers below the detection limit (0.02 µg/L for all compounds) were first replaced  
2 by half the detection limit. Then, concentrations were expressed in molar units by  
3 using the following conversion factors: PCB congeners 138 and 153: 1 µg = 2.771  
4 nmol; PCB congener 180: 1 µg = 2.530 nmol; HCB: 1 µg = 4.02 nmol; and p,p'-DDE:  
5 1 µg = 3.14 nmol. Finally, marker PCBs in molar units were added.

6 Data that were not normally distributed, were ln-transformed (natural logarithm) and  
7 described by their geometric means and 95% confidence intervals (CIs).

8 *Multiple linear regression models of serum PCBs in adolescents and adults.* Multiple  
9 linear regression models were build to identify the major determinants of the  
10 biomarker levels. In first instance, this was done for serum PCBs in adolescents and  
11 adults. For both age groups, separate models were investigated for different units of  
12 PCBs, i.e. (A) PCB concentration expressed in nmol/L without any adjustment for  
13 blood fat (volume-based units); (B) PCB concentration expressed in nmol/L with  
14 blood fat (mg/dL) added as a separate explanatory variable in the model (volume-  
15 based units with adjustment for blood fat); and (C) PCB concentration expressed in  
16 pmol per gram blood fat (fat-based units). Possible covariates that determine inter-  
17 individual variation of the concentration of PCBs were listed based on a literature  
18 search. A multiple linear regression model was build including covariates significant  
19 at a 10% level in the univariate analyses. Important covariates are identified by  
20 stepwise regression procedures in which we set the p-value at 0.10 for the  
21 independent variables to enter and at 0.05 to stay in the model. The adjusted R-  
22 square (coefficient of determination,  $R^2_a$ ) of the obtained multiple linear regression  
23 models was used to select the “best” model. This coefficient shows the proportion of  
24 variability in the biomarker values that is accounted for by this model, penalizing for  
25 the number of explanatory variables in the model.

1 Variance inflation factors (VIFs) were used to analyse the effects of multicollinearity.  
2 If the VIFs were larger than 10 multicollinearity was concluded.  
3 The assumptions of normality, constancy of variance, independence (randomness)  
4 and linearity were checked with informal diagnostic plots and formal tests (White's  
5 General test for constancy of variance, Kolmogorov-Smirnov test for normality and  
6 the lack of fit test for linearity).

7 *Impact of outliers.* An observation is considered to be outlying with respect to its  
8 biomarker value when its value lies more than 1.5 times above or below the  
9 interquartile range (IQ = the 75<sup>th</sup> percentile minus the 25<sup>th</sup> percentile). The impact of  
10 outliers on the summary statistics was investigated. A case may also be outlying if it  
11 has dramatic effects on the multiple regression analysis. A case may be outlying with  
12 respect to its Y value (biomarker value), its X value (covariates in the model), or both.  
13 In a first step, we identified outlying cases based on the fitted model; outliers with  
14 respect to the response were assessed using the studentized deleted residuals and  
15 outliers with respect to the predictor variables were assessed using the leverages or  
16 the diagonal elements of the hat matrix. In a second step, specific methods (DFFITS  
17 and Cook's Distance) were used to detect if these outlying cases were influential,  
18 that is cases that heavily influenced the fitted model. Models were fitted with and  
19 without influential cases and it was studied whether exclusion of the influential cases  
20 significantly changed the regression parameters.

21 *Quantification of impact of determinants.* Quantitative relationships between the  
22 determinants and the biomarkers were calculated from the estimates of the multiple  
23 linear regression model, assuming that, when quantifying the relation of one  
24 covariate with the biomarker, all other covariates in the model are fixed at the  
25 population mean.

## 1 **RESULTS**

2 Descriptive statistics and mean exposure values for the sum of marker PCBs, HCB  
3 and p,p'-DDE in the three study populations are given in Table 1.

4

### 5 *Choice of units: volume-based versus fat-based units*

6 Both in adolescents and in adults, we build multiple linear regression models to  
7 identify the determinants for inter-individual variability of serum PCB levels. These  
8 regression models were constructed for different units of serum PCBs, i.e. serum  
9 PCBs in nmol/L without adjustment for blood fat (model A), serum PCBs in nmol/L  
10 with blood fat as explanatory variable in the model (model B) and serum PCBs  
11 expressed in pmol/g blood fat (model C). In both age groups, inclusion of blood fat in  
12 the analysis, either as an independent variable in the model (B) or as a correction  
13 factor in the formula (C), considerably increased the variability in serum PCB levels  
14 that is explained by the model (Table 2). In adolescents, very similar results were  
15 obtained with blood fat added as an independent variable in the model ( $R^2_a=0.43$ )  
16 and with fat-based units ( $R^2_a=0.44$ ). In adults, inclusion of blood fat as a separate  
17 variable in the model resulted in a considerably higher adjusted  $R^2$  ( $R^2_a=0.30$ )  
18 compared to a model with fat-based units ( $R^2_a=0.19$ ). Thus, we conclude that the  
19 model with PCBs in volume-based units with blood fat as separate independent  
20 variable in the multiple regression model is superior to the other units (i.e. explaining  
21 more variability) and will therefore be used in all further analyses.

22

### 23 *Impact of outliers*

24 An observation may be outlying with respect to its biomarker values. For serum PCBs  
25 in nmol/L in adolescents, 17 cases were found to be beyond the fences. Dropping

1 these outlying cases from the study population resulted in a similar geometric mean  
2 and 95% confidence interval (mean: 0.779 nmol/L; 95% CI: 0.761-0.798 nmol/L)  
3 compared to the whole study population (mean: 0.781 nmol/L; 95% CI: 0.762-0.801  
4 nmol/L). In adults, 37 cases were identified as outlying. Here also, the geometric  
5 mean and confidence interval without the outlying cases (mean: 5.500 nmol/L; 95%  
6 CI: 5.396-5.605 nmol/L) were comparable with those of the whole population (mean:  
7 5.485 nmol/L; 95% CI: 5.370-5.603 nmol/L).

8 A case may also be an outlier if it has dramatic effects on the regression coefficients  
9 in the linear regression models. In adolescents, 57 cases were found to be outlying  
10 with respect to the biomarker values while 55 cases were outliers with respect to the  
11 covariates; respectively 40 and 13 of them were influential. Fitting the model with or  
12 without the 53 influential cases changed the adjusted  $R^2$  of the model from 0.43 to  
13 0.51. In adults, 67 cases were outliers with respect to the biomarker values and 50  
14 cases with respect to the covariates. Respectively 54 and 14 of them were found to  
15 be influential. Exclusion of the 68 influential cases resulted in an adjusted  $R^2$  of 0.38  
16 compared with an adjusted  $R^2$  of 0.30 with influential cases. Thus, identification and  
17 exclusion of influential cases in multiple regression models will lead to better  
18 prediction models, with an higher proportion of the variability explained. Also,  
19 exclusion of influential outliers sometimes changed the regression coefficients (but  
20 not the sign) of the covariates and had an influence on the covariates that were  
21 retained as significant ( $p < 0.05$ ) in the model. When we exclude influential cases,  
22 these models will be better applicable to the general population, and are therefore  
23 preferential for studies in which we want to extrapolate general guidelines for  
24 biomonitoring.

25

1 *Determinants of serum PCBs, HCB and p,p'-DDE.*

2 For all age groups separately, we constructed multiple linear regression models with  
3 PCAHs expressed in nmol/L and without influential outliers in order to identify the  
4 factors that determine inter-individual variability in the serum levels of each of the  
5 three PCAHs. For serum PCBs, the model explained 39.1%, 50.8% and 38.0% of the  
6 variability in respectively newborns, adolescents and adults (Figure 1). For the  
7 chlorinated pesticides, the total variability explained was lower than in PCBs and  
8 amounted in the three age groups to respectively 21.5%, 26.0% and 43.0% for HCB,  
9 and to 22.1%, 27.9% and 26.0% for p,p'-DDE. The factors that contributed most to  
10 the variability were blood fat, age, sex, body-mass index, change in body weight,  
11 area of residence and local meat consumption (Figure 1).

12

13 *Quantification of impact of determinants.*

14 Starting from the multiple linear regression models, we used the estimates of the  
15 regression coefficients to quantify the relationship between the biomarkers of  
16 exposure and significant determinants at the 5% level (Table 3).

17 We found a highly consistent relationship between changes in the blood fat content  
18 and variations in the serum concentration of the three fat-soluble pollutants. In  
19 peripheral serum (adolescents and adults), an increase of the blood fat content with  
20 100 mg/dL was associated with a 11 to 14% increase ( $p < 0.001$ ) in the concentration  
21 of PCBs, HCB or p,p'-DDE. In cord plasma, the variation among the three markers  
22 was higher with increases between 11 and 23% ( $p < 0.001$ ) for an augmentation of the  
23 blood fat content with 100 mg/dL.

24 Both in mothers and adults, there was a strong association between the age of the  
25 respondent and the serum/plasma concentration of all three persistent compounds.

1 Increases in the serum levels of the three PCAHs were highly consistent in 50 to 65  
2 year old adults (+11 to 14% per 5 year increase of age,  $p < 0.001$ ) and more diverse in  
3 18 to 40 year old mothers (+18 to 34% per 5 year increase of age,  $p < 0.001$ ). As the  
4 age range in adolescents was only 2.7 years, it was not surprising to find a weak  
5 relationship between the PCAHs concentration and age; we only found a significant  
6 ( $p < 0.01$ ) association between serum HCB and the adolescents' age. On the other  
7 hand, the serum PCB concentration at the age of 14 to 16 years was significantly  
8 ( $p < 0.01$ ) influenced by the age of the mother at the adolescent's birth. Also, birth  
9 weight of the adolescent was associated with HCB levels and the event of being  
10 breastfed as a baby had a major impact on the serum concentration of all three  
11 PCAHs of the teenager. Compared to adolescents who only received formula, those  
12 who were breastfed had 7%, 26% and 32% higher serum concentrations of  
13 respectively HCB, PCBs and p,p'-DDE (all  $p < 0.001$ ).

14 With respect to sex, we found different effects in the various age groups. In adults,  
15 serum levels of p,p'-DDE and HCB were respectively 28 and 40% ( $p < 0.001$ ) higher in  
16 women compared to men, while serum PCBs did not differ significantly between men  
17 and women. In adolescents, we found 20 to 40% ( $p < 0.001$ ) higher serum levels in  
18 boys compared to girls for all three PCAHs. Also for body-mass index, the results  
19 varied by age. In adults, serum levels of p,p'-DDE and HCB (but not PCB) were  
20 significantly and positively related to BMI, while in adolescents, the serum levels of all  
21 three PCAHs decreased significantly when BMI was rising. In the mothers of the  
22 newborns, we found contradicting results, i.e. BMI was negatively associated with  
23 serum PCBs and positively with serum HCB. For different age groups and PCAHs,  
24 however, we found very consistent relationships between changes in body weight



1 and serum levels of persistent compounds: a weight gain of 10 kg was associated  
2 with a decrease in the serum levels of PCBs, HCB or p,p'-DDE up to 24%.

3 In the mothers of the newborns, no associations were found between the food intake  
4 pattern of the last year before pregnancy and serum levels of PCAHs in cord blood.

5 In adolescents, a positive association was found between serum PCBs, HCB or p,p'-  
6 DDE on the one hand and different sources of animal fat (dairy fat, fish fat and added  
7 fats) on the other hand. The serum level of p,p'-DDE was positively associated with  
8 fresh vegetable consumption in adolescents. In adults between 50 and 65 years old,  
9 a positive association was detected between animal fat intake (eel, mussels, chicken)  
10 and serum PCB levels. Contradictory, both total fat and milk fat were associated  
11 negatively with serum levels of HCB or p,p'-DDE in adults.

12 Over and beyond food intake we studied the effect of locally grown food products on  
13 serum levels of PCAHs. In mothers of newborns, consumption of local dairy products  
14 was associated with higher serum PCB levels. In adolescents, consumption of local  
15 meat was related to higher levels of all three PCAHs. In all age groups, local  
16 vegetable consumption and/or being owner of a vegetable garden was associated  
17 with increased serum levels of the persistent pesticides.

18 Finally, a number of variables show small effects in some age groups and for some  
19 exposure markers. Smoking was identified as confounder in adults for PCBs and for  
20 HCB, but the results were inconsistent and the impact was small. In adolescents, a  
21 higher educational level of the respondent itself was significantly associated with  
22 higher serum levels of HCB; the educational level of the family (highest of mother or  
23 father) was significantly and positively related to serum PCBs of the adolescent. In  
24 adults and mothers, significantly higher values of HCB and p,p'-DDE were found in  
25 the winter season.

## 1 **DISCUSSION**

2 In this study, we determined the major covariates of the serum levels of marker  
3 PCBs, HCB and p,p'-DDE in three different age groups. We found consistent  
4 relationships between the serum concentration of the three fat-soluble pollutants and  
5 blood fat, age, changes in body weight and sources of animal fat in the diet. The  
6 serum levels of the two persistent pesticides were positively associated with local  
7 vegetable consumption. In 14- to 15-year old adolescents, being breastfed as a baby  
8 was an important predictor of serum levels of PCAHs during puberty. The impact of  
9 sex and body-mass index differed by age. We also showed that exclusion of  
10 influential outliers in multiple linear regression models resulted in multiple regression  
11 models that are better applicable to the general population. The regression model  
12 with the pollutant expressed in volume-based units and blood fat as a separate  
13 independent variable was superior compared to models with other units in terms of  
14 adjusted R-square.

15  
16 The levels reported in this study are in line with former measurements of serum  
17 levels of PCBs, HCB and p,p'-DDE in Flanders. In 1999, 200 17-18 year old  
18 adolescents and 200 adult women between the age of 50 and 65 were studied in a  
19 pilot trial for biomonitoring. The geometric mean of the summed PCBs in the  
20 youngsters was 293 pmol/g fat (Staessen et al. 2001), which is considerably higher  
21 compared to the current study (i.e. 178 pmol/g fat). Yet, this is in line with the  
22 expected, since the teenagers in the pilot trial were on average 3 years older and the  
23 levels of PCBs in the environment are declining over time. Studies in breast milk  
24 have shown that levels of PCBs have decreased by 80-90% over a period of 12  
25 years (Van Leeuwen and Malisch 2002; Schade and Heinzow 1998). In the pilot

1 project, median levels in 200 women equaled 1054 pmol/g fat for PCBs, 442 pmol/g  
2 fat for HCB and 2736 pmol/g fat for p,p'-DDE (Koppen et al. 2002). These levels are  
3 also higher than the average values measured in the current study, that is 919, 273  
4 and 2341 pmol/g fat, respectively. Possible explanations for the lower values in the  
5 current trial are the decreasing levels in the environment and the fact that in the pilot  
6 project only women were studied, which may result in higher values for this age  
7 class.

8

9 Up to now, there is no general agreement in literature on the use of units for  
10 biomarkers assessing exposure to persistent pollutants. In some studies,  
11 concentrations (mol/L) are reported, while others use fat-based units (mol/g blood  
12 fat). In this study, we could explain more of the observed variability (highest adjusted  
13 R<sup>2</sup>) if PCAHs were expressed as volume-based units and blood fat was included as  
14 an explanatory variable in the model. Results were only presented for PCBs, but  
15 similar conclusions can be drawn for the chlorinated pesticides, HCB and p,p'-DDE  
16 (data not shown). For all three components, the relation between the serum  
17 concentration and the blood fat content was highly consistent: an increase of the  
18 blood fat content by 100 mg/dL was associated with a 10 to 14% higher  
19 concentration of PCAHs in peripheral serum (both in adolescents and in adults), and  
20 a 11 to 23% higher concentration in cord blood. Therefore, for multiple regression  
21 analysis of PCBs or chlorinated pesticides, we recommend to use the values  
22 expressed as volume-based units (nmol/L) with blood fat added as a separate  
23 independent variable in the model. This approach allows the serum concentrations of  
24 persistent pollutants to be appropriately adjusted for blood fat, and the statistical  
25 significance of other variables in the model to be independent of effects of blood fat

1 concentrations. These recommendations are in line with guidelines for the use of  
2 creatinine in measurements of urinary biomarkers by Barr et al. (Barr et al. 2005).  
3 Their advice for biologic monitoring measurements in spot urine is to use the urinary  
4 analyte concentration and add creatinine as a separate value in the multiple  
5 regression model. Urinary creatinine concentrations are used to adjust for dilution of  
6 spot urine samples, and thus are comparable to the use of blood fat as dilution  
7 marker for non-fasting serum or plasma.

8

9 In most studies, the major determinant of biomarkers that reflect persistent pollutants  
10 is the age of the respondent (Schade and Heinzow 1998; Dallaire et al. 2002; Falk et  
11 al. 1999; Jonsson et al. 2005; Solomon and Weiss 2002; Sweeney et al. 2001). In the  
12 current study, we also found a highly consistent relationship between biomarkers of  
13 exposure and age in 50 to 65 year old adults (+11% for PCBs, +14% for HCB and +  
14 14% for p,p'-DDE per 5 year increase of age), while the effects were less constant in  
15 18 to 40-year old pregnant women (+34%, +18 and +24%, respectively). Possibly,  
16 other factors such as parity or previous breastfeeding, although not identified as  
17 significant determinant, may interact with age in young mothers. The age  
18 dependency can be explained by larger buildup of PCAHs in the body throughout life.  
19 Moreover, as environmental loads of PCAHs are decreasing over time, older people  
20 have been exposed to higher levels in the past.

21

22 In adolescents, serum PCAHs were negatively related to BMI, while a positive  
23 relationship with BMI was observed in adults. These findings are in accordance with  
24 earlier studies in youngsters (Nawrot et al. 2002; Ryan et al. 1994) and in adults  
25 (Schade and Heinzow 1998; Falk et al. 1999; Jonsson et al. 2005). In adults, we

1 expect a positive relationship in view of the higher absolute body burden in obese  
2 subjects and the assumption of a steady state between PCAH concentration in fat  
3 tissue and in serum. In adolescents, a transient dilution effect in adipose tissue  
4 during growth spurt may explain the negative relationship between serum PCAH  
5 levels and BMI. As body fat content in teenager girls is about double that of boys in  
6 adolescence, the same mechanism may explain the lower serum values in girls  
7 compared to boys. Also the negative relationship between changes in body weight  
8 and serum levels of PCAHs is in line with this mechanism. Weight loss and lipolysis  
9 may lead to an up-concentration of PCAHs in the lipid tissue and subsequently, if we  
10 accept the assumption of steady state between fat tissue and serum, to higher serum  
11 values of PCAHs.

12

13 Nutrition is a major source of exposure to PCAHs. In mothers of newborns, we found  
14 no significant associations between serum levels of PCAHs and food consumption in  
15 the year before the pregnancy. Although this best represents the 'usual' food intake,  
16 the relationship between serum levels and food consumption is probably is probably  
17 troubled by the long time gap, by changing food habits during pregnancy and  
18 possibly also by a changing metabolism during pregnancy. In adolescents, the serum  
19 levels of PCAHs at the age of 14 to 16 years were associated with the use of  
20 breastfeeding as baby. This observation has been reported several times in the last  
21 decade and in different Western countries (Schade and Heinzow 1998; Falk et al.  
22 1999; Jonsson et al. 2005; Solomon and Weiss 2002; Nawrot et al. 2002; Glynn et al.  
23 2007; Grimvall et al. 1997). Still, according to WHO guidelines, mother's milk remains  
24 the best nutrition for a baby (WHO 2001). With some exceptions, we can state that  
25 both in adolescents and in adults, serum levels of PCAHs were positively related to

1 animal food intake (meat, chicken or fish). This is in accordance with literature (Falk  
2 et al. 1999; Glynn et al. 2007; Glynn et al. 2003; He et al. 2001) and with the fact that  
3 animal fats are the major source of fat-soluble bio-accumulating pollutants. The  
4 serum levels of the chlorinated pesticides (but not PCBs) were positively associated  
5 with local vegetable consumption and/or with being owner of a vegetable garden.  
6 This indicates that there may be local contamination, either as a result of former use  
7 of DDT and HCB in the area, or even due to current illegal or ignorant use of these  
8 products in Flanders. As a result of the current biomonitoring study, the government  
9 has launched a campaign to collect old pesticides through the local community  
10 authorities.

11

12 Finally, a number of variables that are associated with the serum levels of PCAHs,  
13 are probably no determining factor in itself but rather a dummy for other underlying  
14 factors that determine exposure to or metabolism of the persistent compounds.  
15 Smoking was identified as confounder in adults, but the results were inconsistent and  
16 the impact was small. Possibly, there is co-linearity between smoking and other life  
17 style factors such as nutrition, or social class. In adolescents, a higher educational  
18 level of the respondent itself was significantly associated with higher serum levels of  
19 HCB; the educational level of mother or father was significantly and positively related  
20 to serum PCBs. It is likely that educational level is associated with nutritional habits  
21 (e.g. fish or milk consumption), living conditions, or other life style factors that may  
22 influence exposure to persistent compounds. The serum concentrations of HCB and  
23 p,p'-DDE differed by season. In adults and mothers, significantly higher values were  
24 found in the winter season. We can hypothesize that this may be due to nutritional

1 factors (e.g. higher animal fat consumption in winter) or to differences in body  
2 composition (e.g. weight gain in winter).

3

4 A major strength of this study was the availability of a large data set (more than 4400  
5 individuals) for three different biomarkers of the same chemical class in different age  
6 classes (from newborn to adults). The chemical measurements, the collection of the  
7 questionnaire data, the calculation of secondary variables and the statistical analyses  
8 were performed in a consistent and uniform method for the three biomarkers and the  
9 three age groups. This has resulted in high quality data and thus allows to generalize  
10 the findings and deduct a number of recommendations for biomonitoring studies on  
11 how to handle outliers, use of units and selection of major confounders. The data  
12 show that methodological aspects and host factors are important to take into account  
13 in biomonitoring studies that perform between-group comparisons or describe  
14 relations between internal PCAH exposure and health effects.

15

16 Although the results are based on a large data set, the conclusions are still limited to  
17 the three biomarkers at exposure levels that were assessed in this study and to the  
18 age ranges of the subgroups that were included in the study. To further improve our  
19 insight in the determinants of variability in these and other biomarkers and the  
20 methodological aspects of biomonitoring studies, we need well-designed studies with  
21 large and well-documented gradients in environmental, biological and life style  
22 factors. A European biomonitoring project, enabling comparison of biomarker data  
23 and information on covariates of citizens of different countries could meet these  
24 requirements.

25

1

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3

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6

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10 content.

11



1 **ABBREVIATIONS**

2

3 95% CI: 95% confidence interval

4 AMAP: arctic monitoring and assessment programme

5 BMI: body-mass index

6 FFQ: food frequency questionnaire

7 HCB: hexachlorobenzene

8 IQ: interquartile range

9 ln: natural logarithm

10 Na<sub>2</sub>-EDTA: sodium ethylenediaminetetraacetic acid

11 ns: not significant

12 PCAHs: polychlorinated aromatic hydrocarbons

13 PCBs: polychlorinated biphenyls

14 p,p'-DDE: p,p'-dichlorodipenyldichloroethylene

15 R<sup>2</sup><sub>a</sub>: adjusted coefficient of determination

16 VIF: variance inflation factor

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**Table 1.** Descriptive statistics in three age groups.

		<b>Mothers &amp; newborns</b>	<b>Adolescents</b>	<b>Adults</b>
N		1169	1679	1583
Sex	female / male	1196 / 0	792 / 887	808 / 775
Age (years)	mean $\pm$ SD	29.6 $\pm$ 4.3	14.9 $\pm$ 0.5	57.6 $\pm$ 4.1
	range	18.1 - 44.0	13.8 -16.5	49.8 - 65.3
BMI	mean $\pm$ SD	23.3 $\pm$ 4.2	20.6 $\pm$ 3.1	26.9 $\pm$ 4.2
	range	14.0 – 44.6	13.7 – 36.6	15.1 – 48.1
Smoking	% non-smokers	64.2%	86.4%	44.6%
	% former smokers	19.6%	-	37.2%
	% current smokers	16.2%	13.6%	18.3%
Blood fat in mg/dL	mean	209	444	612
	(95% CI)	(205-213)	(440-448)	(604-620)
Serum PCBs in nmol/L	geometric mean	0.333	0.781	5.485
	(95% CI)	(0.318-0.349)	(0.762-0.801)	(5.370-5.603)
Serum PCBs in pmol/g fat	geometric mean	226	178	919
	(95% CI)	(215-236)	(174-183)	(901-937)
Serum HCB in nmol/L	geometric mean	0.189	0.364	1.370
	(95% CI)	(0.180-0.198)	(0.358-0.370)	(1.329-1.412)
Serum HCB in pmol/g fat	geometric mean	101	87.9	273
	(95% CI)	(95-106)	(86.4-89.4)	(264-281)
Serum p,p'-DDE in nmol/L	geometric mean	0.993	1.462	8.885
	(95% CI)	(0.929-1.056)	(1.404-1.522)	(8.441-9.352)
Serum p,p'-DDE in pmol/g fat	geometric mean	515	529	2341
	(95% CI)	(481-548)	(486-572)	(2202-2479)

PCBs: sum of marker PCB 138, 153 and 180; HCB: hexachlorobenzene; p,p'-DDE: p,p'-dichlorodiphenyldichloroethane (metabolite of DDT).

**Table 2.** Adjusted R-square of multiple linear regression models with different units of serum PCBs as dependent variable.

	Adolescents	Adults
<b>Adjusted R<sup>2</sup> of multiple linear regression models*</b>		
(A) volume-based units of serum PCBs	R <sup>2</sup> <sub>a</sub> = 0.3860	R <sup>2</sup> <sub>a</sub> = 0.1602
(B) volume-based units of serum PCBs with adjustment for blood fat	R <sup>2</sup> <sub>a</sub> = 0.4281	R <sup>2</sup> <sub>a</sub> = 0.2990
(C) fat-based units of serum PCBs	R <sup>2</sup> <sub>a</sub> = 0.4443	R <sup>2</sup> <sub>a</sub> = 0.1862

\* the multiple regression model was adjusted for the following covariates: *in adolescents*: BMI, sex, educational level of the family, breastfeeding, smoking during pregnancy of the mother, age mother at childbirth, study area, local meat consumption, eel fat consumption, animal fat consumption and milk fat consumption (only for A); *in adults*: age, smoking, change in body weight, BMI (only for C), study area, local meat consumption (only for B and C), eel fat consumption, mussel fat consumption, chicken fat consumption, vegetable fat consumption (only for C), consumption of liver fat (only for A) and alcohol use (only for C).

**Table 3.** Quantification of determinants of serum PCBs, hexachlorobenzene (HCB) and p,p'-DDE.

	Mothers & newborns cord plasma			Adolescents peripheral serum			Adults peripheral serum			
	PCBs	HCB	p,p'-DDE	PCBs	HCB	p,p'-DDE	PCBs	HCB	p,p'-DDE	
N	1011	1003	1094	1387	1390	1493	1438	1506	1468	
<b>% change in biomarker concentration for a given change of the covariate</b>										
<b><i>covariates related to blood sampling</i></b>										
blood fat	+ 100 mg/dL	+23%***	+11%***	+20%***	+14%***	+11%***	+11%***	+11%***	+10%***	+12%***
season	spring vs. winter	ns	-20%***	-23%***	ns	ns	ns	ns	-17%***	-14%**
	summer vs. winter	ns	-18%**	-18%**	ns	ns	ns	-	-	-
	autumn vs. winter	ns	-19%**	ns	ns	ns	ns	ns	ns	ns
<b><i>biological covariates</i></b>										
age	+5 years	+34%***	+18%***	+24%***	ns	+5%*** <sup>a</sup>	ns	+11%***	+14%***	+14%***
sex	male vs. female	-	-	-	+40%***	+20%***	+26%***	ns	-40%***	-28%***
BMI	+5 kg/m <sup>2</sup>	-15%***	+6%*	ns	-40%***	-14%***	-24%***	ns	+24%***	+22%***
weight change	+10 kg	ns	ns	ns	ns	-7%*** <sup>b</sup>	ns	-14%***	-19%***	-24%***
birth weight	+1 kg	-	-	-	ns	+3%*	ns	-	-	-
age of mother at birth	+5 years	-	-	-	+3%**	ns	ns	-	-	-
<b><i>socio-economic covariates</i></b>										
highest educational level of the family (mother or father)	primary vs. higher	ns	ns	ns	-19%*	ns	ns	ns	ns	ns
	low secondary vs. higher high secondary vs. higher				-10%***					
secondary school level of the adolescent	technical vs. general	-	-	-	ns	-4%*	ns	-	-	-
	vocational vs. general					-7%**				

-: not applicable (information not available or not relevant); ns: not significant. \* p < 0.05; \*\* p < 0.01; \*\*\* p < 0.001. § added fats = spreading and cooking fats

<sup>a</sup> for increase of age with 1 year; <sup>b</sup> for weight change of 5 kg

Quantitative relationships between a covariate and a biomarker were calculated from the estimates of the multiple linear regression model, assuming that, all other covariates in the model are fixed at the population mean.

**Table 3, continued.** Quantification of determinants of serum PCBs, hexachlorobenzene (HCB) and p,p'-DDE.

		Mothers & newborns			Adolescents			Adults		
		cord plasma			peripheral serum			peripheral serum		
		PCBs	HCB	p,p'-DDE	PCBs	HCB	p,p'-DDE	PCBs	HCB	p,p'-DDE
N		1011	1003	1094	1387	1390	1493	1438	1506	1468
<b>% change in biomarker concentration for a given change of the covariate</b>										
<b>life style</b>										
smoking	former vs. non-smoker	ns	ns	ns	-	-	-	ns	-6%*	ns
	current vs. non-smoker	ns	ns	ns	ns	ns	ns	+10%***	ns	ns
breastfed as baby	yes vs. no	-	-	-	+26%***	+7%***	+32%***	-	-	-
<b>food intake</b>										
total fat	+10 g/day	ns	ns	ns	ns	ns	ns	ns	ns	-10%**
milk fat	+5 g/day	ns	ns	ns	+1.3%**	+4%**	ns	ns	-7%**	-14%**
fish fat (all fish)	+5 g/day	ns	ns	ns	ns	+6%*	ns	ns	ns	ns
eel fat	+0.1 g/day	ns	ns	ns	+1.0%*	ns	ns	+1.3%***	ns	ns
mussel fat	+0.1 g/day	ns	ns	ns	ns	ns	ns	+4%**	ns	ns
chicken fat	+0.1 g/day	ns	ns	ns	ns	ns	ns	+1%*	ns	ns
added fats§	+5 g/day	ns	ns	ns	ns	+1.3%**	ns	ns	ns	ns
fresh vegetables	+100 g/day	ns	ns	ns	ns	ns	+2%*	ns	ns	ns
<b>local food consumption</b>										
local meat	yes vs. no	ns	ns	ns	+16%***	+7%**	+19%***	+10%***	ns	ns
local dairy	yes vs. no	+11%*	ns	ns	ns	ns	ns	ns	ns	ns
local vegetables	yes vs. no	ns	+12%*	+18%**	ns	+6%**	+13%*	ns	ns	ns
vegetable garden	yes vs. no	ns	ns	ns	ns	ns	+14%***	ns	ns	+11%*

-: not applicable (information not available or not relevant); ns: not significant. \* p < 0.05; \*\* p < 0.01; \*\*\* p < 0.001. § added fats = spreading and cooking fats

<sup>a</sup> for increase of age with 1 year; <sup>b</sup> for weight change of 5 kg

Quantitative relationships between a covariate and a biomarker were calculated from the estimates of the multiple linear regression model, assuming that, all other covariates in the model are fixed at the population mean.

**Figure 1.** Adjusted R-square representing the percent variability explained by the covariates in the multiple regression models for serum PCBs, hexachlorobenzene (HCB) and p,p'-DDE in three age classes.

